Journal of the Marine Biological Association of the United Kingdom

cambridge.org/mbi

Research Article

Cite this article: Freitas D, Arenas F, Vale CG, Pinto IS, Borges D (2023). Warning of warming limpets: sea temperature effects upon intertidal rocky assemblages. *Journal of the Marine Biological Association of the United Kingdom* **103**, e7, 1–9. https://doi.org/10.1017/ S0025315422001096

Received: 23 February 2022 Revised: 1 October 2022 Accepted: 17 November 2022

Key words:

Activation energy; metabolic rate; metabolic theory of ecology; oxygen consumption tests; Patella depressa; Patella rustica; Patella ulyssiponensis; Patella vulgata

Author for correspondence: Débora Borges, E-mail: debora.borges@ciimar.up.pt

© The Author(s), 2023. Published by Cambridge University Press on behalf of Marine Biological Association of the United Kingdom



Warning of warming limpets: sea temperature effects upon intertidal rocky assemblages

Diana Freitas¹, Francisco Arenas², Cândida Gomes Vale², Isabel Sousa Pinto^{1,2} and Débora Borges²

¹Department of Biology, Faculty of Sciences, University of Porto, Porto, Portugal and ²Interdisciplinary Centre of Marine and Environmental Research, University of Porto, Porto, Portugal

Abstract

Limpets (Patella spp.) are marine gastropods that inhabit rocky shores along the coasts of Europe, the Mediterranean, Macaronesia and the north-west coast of Africa. Being considered key species, limpets have an important role regulating algal assemblages in coastal communities. The goal of this work was to evaluate the influence of sea temperature on the respiration rate of four limpet species occurring in mainland Portugal, in line with predictions from the metabolic theory of ecology. The individuals were collected from rocky shores in Portugal and exposed to sea temperatures ranging from 6-28°C for respiration rate assessments. Following the estimation of the relationship between oxygen consumption and temperature the activation energy was calculated. In parallel, low and high thermal thresholds were determined for three of the species. The results indicated that P. ulyssiponensis oxygen consumption increased linearly with sea temperature and the remaining species presented the same tendency. The values of activation energy ranged between 0.33-0.76 eV. For P. ulyssiponensis, the highest activation energy indicated that this species is more sensitive to temperature variations while for the tested temperatures it presented a higher thermal tolerance limit than the other species. Such findings indicate that P. ulyssiponensis is the most susceptible of these species to climate change, in line with the tolerance-plasticity trade-off hypothesis. This work provides a good starting point for understanding the effect of sea temperature on oxygen consumption in Patella spp. and for comprehending sensitivity of limpets to temperature increases under future climate change scenarios.

Introduction

The global climate has changed throughout time and there is evidence that these changes have impacts on organisms and ecosystems (Hoegh-Guldberg & Bruno, 2010). As global temperatures continue to rise, the relative abundance and composition of species continues to change. With projected future climate changes, more dramatic shifts are expected in marine ecosystems (Lawler, 2009; Magris et al., 2014; Lathlean et al., 2017). In the intertidal zone, the organisms are subject to considerable natural environmental variability (Tait & Dipper, 1998) and exposed to several stressors such as water temperature, desiccation and solar radiation, affecting their survival (Seabra et al., 2011, 2016). Other factors such as the micro-topography, air temperature, wind speed, solar radiation and relative humidity also play an important role, particularly in the high shore (Helmuth et al., 2011; Seabra et al., 2011; Lima et al., 2016). These variations in the environment impact intertidal species in multiple ways, affecting their physiological and biochemical processes (Hawkins et al., 2009; Prusina et al., 2014), morphology (Harley et al., 2009), behaviour and biotic interactions (Firth & Crowe, 2010), and ultimately their geographic distributions (Sousa et al., 2012). Temperature variation in particular has a strong influence over organisms, controlling all life stages and conditioning fundamental biological processes, such as the metabolism, growth, reproduction and survival of the individuals (Helmuth et al., 2006; Firth & Crowe, 2010; Moore et al., 2011; Stillman, 2019; Mieszkowska et al., 2021).

Metabolism is a fundamental process that regulates the flux of energy and materials between the organism and the environment (Gillooly *et al.*, 2001). The metabolic theory of ecology (MTE) offers a predictive framework for assessing and understanding changes in the abundance, distribution and diversity of organisms and the fluxes of energy and materials in ecological systems (Brown *et al.*, 2004). Basic metabolic theory predicts that warming might increase the performance of an organism by increasing its metabolic rate, until it reaches a given threshold (Mertens *et al.*, 2015). However, the heterotrophic respiration and autotrophic productivity are scaled differently from temperature (Gillooly *et al.*, 2001; Allen *et al.*, 2005; Mertens *et al.*, 2015). Hence, the MTE suggests that producers will show a weaker temperature response than consumers (Brown *et al.*, 2004; Yvon-Durocher *et al.*, 2010; Mertens *et al.*, 2015). The MTE predicts that grazers will be more sensitive to warming than their food sources. In fact, it has been used to predict increases in top-down control of algal assemblages (O'Connor, 2009) responsible for surpassing the compensatory/grazing metabolic threshold. Afterwards, the community shifts from a top-down (herbivore-regulated) to a bottom-up (algae-regulated) system (Hawkins *et al.*, 2009; O'Connor, 2009; O'Connor *et al.*, 2011). The effect of reducing or removing grazers on the community is known to result in important changes in the community structure (Coleman *et al.*, 2006).

Limpets promote spatial variation at northern latitudes by preventing establishment of fucoid algae, which colonize the exclusion plots, and decrease variation in the south, where outcomes are more contingent (Coleman et al., 2006). Limpets are mainly herbivorous marine gastropods that inhabit exposed and sheltered rocky shores along the western coast of Europe, the Mediterranean coast, the north-west of Africa and the Macaronesian islands (Guerra & Gaudencio, 1986; Borges et al., 2015). They are considered key species that play an important ecological role in marine communities, exerting strong competitive intraspecific and interspecific pressures (Underwood, 1979; Underwood & Jernakoff, 1981; Boaventura et al., 2002). As grazers, limpets feed on large amounts of seaweed propagules, controlling their growth and regulating the ecological succession of biological communities established in the intertidal areas (Boaventura et al., 2002). Hence, limpets are excellent models to investigate the MTE grazers-algal relationships described above.

The Portuguese coast constitutes a biogeographic transition zone (from the Mediterranean to the Atlantic biogeographic regions), resulting in high environmental variability with effects on populations dynamics (Boaventura et al., 2002; Pereira et al., 2006; Spalding et al., 2007). In the north-west coast of Portugal, the most abundant species are Patella ulyssiponensis Gmelin, 1961, Patella vulgata Linnaeus, 1758 and Patella depressa Pennant, 1777 (Casal et al., 2018). These three species coexist in the intertidal zone but at different tide levels. Usually, P. ulyssiponensis occurs on the low shore and can also be found in tide pools on the mid-shore (Boaventura et al., 2002). Patella vulgata occupies mainly the region between the mid shore and the upper shore (Blackmore, 1969; Boaventura et al., 2002). Patella depressa is present from the mid shore to the low shore, coexisting with P. vulgata (Boaventura et al., 2002; Silva et al., 2003). Patella rustica Linnaeus, 1758 is more abundant in the southern coast of Portugal than the north-western coast (Lima et al., 2006; Ribeiro et al., 2009). This species occupies the upper shore in the intertidal zone (Lima et al., 2006). These species have overlapping distributions: P. ulyssiponensis, P. vulgata and P. depressa can be found from the British Isles to Portugal (Boaventura et al., 2002; Borges et al., 2015). Patella ulyssiponensis and P. depressa are also present in North Africa (Borges et al., 2015). Patella rustica can be found from the Mediterranean Sea to the Atlantic coast of the Iberian Peninsula and North Africa (Lima et al., 2006). In this context, the goal of this study was to evaluate the influence of sea temperature on the metabolic rate of four species of the genus Patella present along the coast of Portugal. Thus, it contributes to the ability to understand and forecast climate change impacts upon limpet populations and algal assemblages across Portuguese continental seashores. The tested hypothesis was that increases in seawater temperature will result in increases in the metabolic rate across the four species of *Patella* in line with predictions from the MTE. For this purpose, trials were performed for assessing oxygen consumption rates in adult limpets under different sea temperatures. The experiment included water temperatures normally experienced along the Portuguese coast (e.g. 12 and 16°C), but also extreme temperatures (e.g. 6 and 28°C). Additionally, this work intended to study the effects of seawater temperature on the limpet probability of survival for three species of the genus Patella and to obtain their maximum and minimum temperature limits of survival.

Materials and methods

Sampling sites

The organisms were collected in the intertidal zone during low tide at three rocky shores in the north of Portugal: Viana

(Viana do Castelo: 41.696944°N 8.853056°W); Molhe (Porto: 41.157583°N 8.683500°W); Aguda (Vila Nova de Gaia: 41.045333°N 8.653139°W); and one in the centre of Portugal: Barcos (Peniche: 39.376556°N 9.339947°W) (Figure 1). The limpets were haphazardly collected from the rock with a pocketknife and immediately transported to the laboratory in net bags inside small coolboxes filled with seawater. Their size in length ranged from 26.98 ± 2.56 mm for *Patella rustica*, 32.05 ± 3.70 mm for *Patella depressa*, 33.88 ± 5.02 mm for *Patella ulyssiponensis* and 34.93 ± 3.29 mm for *Patella vulgata*. In addition, some stones with biofilm were also collected for food provision during the acclimation period.

Acclimation period and routine procedures

In the laboratory, each species was exposed to two acclimation periods. During the first period, individuals were placed on the top of rocks for one week, inside tanks with aeration and filtered seawater at the same sea temperature as the collection site. This period allowed the species to adjust to new conditions. After that period, the specimens were moved to a new system. This system comprised 14 tanks, grouped in pairs used as water baths under different temperatures (6, 8, 12, 16, 20, 24 and 28°C). Each tank had two cylinders with 2 litres capacity, and inside each cylinder one rock was placed with two limpets on top and covered by a mesh to prevent their escape (Figure 2). The tanks were aerated, and the seawater flow renewed every 20 min where new seawater entered the cylinder for 1 min. Overall, for each temperature tested and species there were a total of eight limpet replicates. The water temperature in each tank was increased or decreased by 1°C per day from the initial temperature of collection up to the established temperature to be tested. When the target temperature was achieved, limpets were left under these settings for 2 days for adaptation to the new conditions until the metabolic rate trials started.

Every day, during both acclimation periods, the mortality and temperature in each tank were recorded. The limpets that died were removed from the system.

Metabolic rate

The metabolic rate estimated was the oxygen consumption or respiration rate of 56 individuals per species. The system for the metabolic trials consisted of three simultaneous and independent chambers with one limpet placed inside each chamber, on top of a rock, with a known volume of filtered seawater. Prior to measuring oxygen consumption under different temperatures, the shells were cleaned to remove any attached organisms. To maintain the established temperature of the water in each chamber, a peristaltic pump transferred the water bath over the chambers. The incubation period lasted for 1 h and 20 minutes, the oxygen concentrations were measured every 20 s with optical O_2 sensors (PreSens Oxy-4) and recorded with the aid of software PreSens Measurement Studio 2.

For each temperature, the oxygen consumption per individual was calculated from oxygen concentration changes in each chamber during the incubation corrected by the volume of water inside each chamber. During the trials, after the stabilization period, the oxygen consumption varied negatively with time. The absolute value of the slope for the initial linear trend represents the oxygen consumption under the tested temperature. This slope was extracted from all replicates, the blank controls were subtracted, and the mean slopes were calculated per temperature. The mean values were plotted to assess the relationship between oxygen consumption and temperature. Afterwards, the shells were removed, and the limpet body mass was left for 48 h at 60°C to



Fig. 1. Location of the sampling sites along the coast of Portugal.

estimate the Dry Weight (DW). The respiration rate was expressed as μ mol O₂ g $_{DW}^{-1}$ h⁻¹.

Activation energy

The metabolic theory of ecology (MTE) is based on the principle that metabolic rate varies according to the organism's body mass and the environmental temperature, as described by equation (1) (Brown *et al.*, 2004):

$$I = I_a M^{(3/4)} e^{-E_a/kT}$$
(1)

where *I* is the metabolic rate, I_o is a normalization constant, independent of the body size and temperature, *M* is the body mass (g) raised to 3/4 power, E_a is the activation energy expressed in electron

volts (eV), *K* is Boltzmann's constant $(8.61733 \times 10^{-5} \text{ eV K}^{-1})$ and *T* is the absolute temperature (kelvin).

The terms of equation (1) can be rearranged to yield equation (2). This equation considers 'mass-corrected' metabolic rate, I, by incorporating the logarithm of mass raised to the 3/4 power.

$$L_n(IM^{-3/4}) = -E_a(1/kT) + L_n(I_o)$$
(2)

This method facilitates quantitative evaluation of the mass and temperature dependence predicted by equation (1), by incorporating the predicted scaling into the analysis and into the y-axis of bivariate plots (Brown *et al.*, 2004). The slope of this relationship gives the activation energy of metabolism. The activation energy (*E*) reflects the sensitivity of the metabolic rate to temperature (Watson *et al.*, 2014). Therefore, following the estimation of



Fig. 2. Scheme of the second acclimation period: system with 14 tanks, grouped in pairs used as water baths under different temperatures.

relationship between oxygen consumption and temperature, for each species, the respective E was calculated.

Mortality trials

In parallel with the assessment of the influence of temperature on the oxygen consumption, the relation of temperature to limpet probability of survival was investigated for *P. ulyssiponensis*, *P. vulgata* and *P. depressa*, but not for *P. rustica* due to the reduced number of specimens available, since a total of 84 individuals were used per species. In these trials, during the second acclimation period, the water temperature was increased or decreased by 1°C every 30 minutes, until the final target temperature was achieved (6, 8, 12, 16, 20, 24 and 28°C). For each temperature, there were two cylinders, each with three limpets placed on the top of a rock, making up a total of 12 replicates per species. The limpets were then left in the tanks at the final temperature and the ratio of alive/dead individuals was estimated every day for 15 days. Mortality was assessed by probing limpets that were detached from the rock and when no reaction was shown, individuals were considered dead. Then, mortality was calculated as the sum of all dead individuals after 15 days of exposure to each temperature. To ensure starvation was not influencing the probability of survival records, the rocks in each tank were replaced every 5 days.

For each species, the low thermal threshold for 50% of the population (LTT_{50}) and the high thermal threshold for 50% of the population (HTT_{50}) (Sagarin *et al.*, 1999; Tang *et al.*, 2000) were calculated based on the probability of survival plots for lower (6–12°C) and higher (16–28°C) temperatures, respectively.

Results

Metabolic rate

The oxygen consumption rates (μ mol O₂ g $_{DW}^{-1}$ h⁻¹) seemed to rise with the increasing temperatures (°C), for the four species (Figure 3). The oxygen consumption rate of *P. ulyssiponensis*



Fig. 3. Mean oxygen consumption (N = 6) (μ mol O₂ g $_{DW}^{-1}$ h⁻¹) variation with temperature (°C) for each limpet species.



Temperature (1/kT)

Fig. 4. Relationship between mass-corrected metabolic rate (L_n (IM^{-3/4})) and sea temperature (1/kT) for each individual of: (A) P. ulyssiponensis and (B) P. vulgata.

significantly increased with temperature (P < 0.05) (Figure 3A). The same pattern was observed for the remaining species, although with no statistical significance (P > 0.05).

Activation energy

The activation energy values obtained for the four species ranged between 0.33–0.76 eV (Figures 4 and 5). *Patella ulyssiponensis* presented the highest activation energy value (0.76 eV). For the remaining three species, the activation energy values were similar, *P. rustica* having the lowest value (0.33 eV). Despite the low r-squared, the *P*-values (P < 0.05) indicate a statistically significant relationship between the mass-corrected metabolic rate (L_n (IM^{-3/4})) and temperature (1/kT).

Mortality trials

The logistic regressions of the probability of survival (%) under lower temperatures (6–12°C) and higher temperatures (16–28°C) were statistically significant (P < 0.05) (Figure 6) for the three species. The LTT₅₀ and HTT₅₀ values were very similar for both *P. vulgata* (10.3 and 24.3°C, respectively) and *P. depressa* (10.7 and 24.8°C, respectively). *Patella ulyssiponensis* had the lower LTT₅₀ value (8.4°C) and a higher HTT₅₀ value (25.3°C) compared with the other species.

Discussion

In the intertidal zone, organisms' metabolism is mostly regulated by the interplay among environmental conditions (e.g. temperature, salinity) and species eco-physiological traits (e.g. body size, activity level, growth and reproduction) (see classical studies from Davies, 1966, 1967; Newell & Roy, 1973; Branch & Newell, 1978; Newell & Branch, 1980; Branch *et al.*, 1988). Although several environmental parameters should be considered, in metabolic rate studies (e.g. respiration rates), temperature is one of the most influencing factors on the organisms' performance (Somero, 2002). Understanding how various species respond to increasing temperatures is crucial to predict the effect of climate change events on natural populations and community structure (Helmuth *et al.*, 2006).

The present study aimed to evaluate the influence of sea temperature on the metabolic rate (in this case oxygen consumption) of four species of the genus *Patella*. The respiration rate of *Patella* spp. was investigated for the first time for populations from the Portuguese coast. The oxygen consumption for *P. ulyssiponensis* increased linearly with temperature within the range of tested temperatures (6–28°C) (Figure 3). The results observed are in agreement with the metabolic theory of ecology (MTE), which predicts that biological rates (e.g. respiration rate) show a similar temperature dependence (Brown *et al.*, 2004). Davies (1966, 1967) studied the effect of body size, temperature and environmental



Fig. 5. Relationship between mass-corrected metabolic rate (L_n (IM^{-3/4})) and sea temperature (1/k7) for each individual of: (A) P. depressa and (B) P. rustica.

acclimation on the metabolic rate of *Patella aspera* (currently *P. ulyssiponensis*) and *P. vulgata*. The results suggested that the low respiration rate of high-level limpets could result from either acclimation to the higher temperatures or lower nutritional conditions of the habitat (Davies, 1967). For the remaining species, *P. vulgata*, *P. depressa* and *P. rustica*, no statistically significant linear relation between the variables was obtained. Despite being non-significant, the results suggest an increase of the oxygen consumption with the increase in temperature (Figure 3), following the tendency observed for *P. ulyssiponensis*. However, to confirm this pattern further studies are required, such as the use of a high number of individual replicates, which will improve the estimation of the relationship between oxygen consumption and temperature.

Activation energy

The activation energy determines the sensitivity of the metabolic rate to temperature (Watson *et al.*, 2014). In this study, the activation energy was determined using oxygen consumption (μ mol O₂ g $_{\rm DW}^{-1}$ h⁻¹) under different temperatures. Gillooly *et al.* (2001) and Brown *et al.* (2004) studied the activation energy of both marine and freshwater species, achieving a general interspecific mean activation energy of 0.65 eV. However, variations from this value are often found ranging from 0.2–1.2 eV (Gillooly *et al.*, 2001). The values of activation energy for *Patella* spp. ranged from 0.33 eV

to a maximum of 0.76 eV, with an average of 0.47 eV (Figures 4 and 5), being within the predicted range by the MTE.

Thermal sensitivity in marine invertebrates varies considerably depending on the taxa. However, it is generally low under highly variable environments, with frequent extreme temperature events, such as those experienced in intertidal habitats (Folguera et al., 2011; Watson et al., 2014). The activation energy obtained for P. ulyssiponensis (0.76 eV) suggests that this species is more sensitive to temperature variations. This could be related to its location in rocky shores, usually found in the low shore or tide pools. Due to this, during the low tides, P. ulyssiponensis is exposed to shorter periods of emersion than the other *Patella* species being protected from extreme temperature variations. For the remaining species, the obtained activation energy values - P. vulgata (0.43 eV), P. depressa (0.36 eV) and P. rustica (0.33 eV) - suggest lower sensitivity of these species to temperature changes. These species are usually found in the mid and upper shore (Boaventura et al., 2002), suggesting a higher capacity to tolerate higher temperatures than species located in the low shore. Such relationship between the thermal sensitivity of the oxygen consumption and the shore level location of the organisms demands further investigation, with intraspecific studies at different tide levels, i.e. individuals of the same species would be collected from the low to high shore and the activation energy for the oxygen consumption investigated at multiple shores and different levels to assess the existence of possible patterns.



Fig. 6. (A) Probability of survival of three species at the end of the experimental period for lower temperatures: (1) *P. ulyssiponensis* ($y = e^{-6-6+0.784x}/1 + e^{-6-6+0.784x}$, P < 0.05); (3) *P. vulgata* ($y = e^{-3.14+0.306x}/1 + e^{-3.14+0.306x}$, P < 0.05); (5) *P. depressa* ($y = e^{-4.29+0.402x}/1 + e^{-4.29+0.402x}$, P < 0.05). (B) Probability of survival of three species at the end of the experimental period for high temperatures: (2) *P. ulyssiponensis* ($y = e^{3.97+0.172x}/1 + e^{3.97+0.172x}$, P < 0.05); (4) *P. vulgata* ($y = e^{7.85+0.323x}/1 + e^{7.85+0.323}/1 +$

Mortality trials

In the intertidal zone, the survival of the organisms living at the interface of the marine and terrestrial realms requires withstanding a broad array of abiotic factors. Temperature gradients depend not only on the climate and seasons, but also fluctuate regularly during emersion and immersion through each tidal cycle (Harley *et al.*, 2009). Understanding the organism's thermal limits will be useful to investigate how the warming climate can influence the behaviour and distribution ranges.

Patella spp. are sensitive to fluctuations in the sea surface temperature, depending on microhabitat topography to oscillations in aerial conditions during emersion time (Chapperon et al., 2016; Lima et al., 2016). Therefore, they are considered indicators of climatic variability (Lima et al., 2006). Particularly, for P. vulgata thermal stress levels are directly linked to increase in water temperature, while high air temperature is only stressful if water temperature is also high (Seabra et al., 2016). For sea temperatures between 6-12°C, the LTT₅₀ of survival was 8.4°C for P. ulyssiponensis, 10.3°C for P. vulgata and 10.7°C for P. depressa. These values are the lower temperature limits for each species studied, indicating that P. ulyssiponensis tolerates lower temperatures than P. vulgata and P. depressa. For a higher range of temperatures (16-28°C), the HTT₅₀ of survival was 25.3°C for P. ulyssiponensis, 24.3°C for P. vulgata and 24.8°C for P. depressa. Therefore, P. ulyssiponensis presented a higher thermal limit than the other two species, P. vulgata and P. depressa. For the same range of temperatures, Patella ulyssiponensis showed a higher thermal tolerance range, which could potentially be linked to their biological features and adaptation to adverse abiotic conditions. According to Evans (1948), limpets from the British Isles can survive temperatures higher than 40°C. Evans (1948) presented a lethal temperature of 42.8°C for P. vulgata, 43.3°C for P. depressa and 41.7°C for P. athelica (now called P. ulyssiponensis). In fact, this semblance in the upper thermal limit for P. vulgata and P. depressa was also reported by Vinagre et al. (2019) when investigating the critical thermal maximum (CTMax) for several coastal taxonomic groups. In such study, P. vulgata presented not only the highest CTMax, but also the highest safety margin among the temperate species studied, with P. depressa values immediately below (Vinagre et al., 2019). Although the present study was only carried out for temperatures ranging between 6-28°C, in future studies, a wider range of temperatures should be considered to fully understand the thermal limits for survival of these species in the Portuguese mainland.

Patella ulyssiponensis respiration rate showed a marked response to temperature and given its highest activation energy is more sensitive to temperature than the other studied species, despite having the highest thermal limit. These results are in line with the tolerance-plasticity trade-off hypothesis, which predicts that individuals already adapted to high temperatures have limited potential to improve their heat tolerance via phenotypic plasticity (Heerwaarden & Kellermann, 2020). Negative relationships between heat tolerance and plasticity have been detected across other tide pool species (see Vinagre et al., 2018), and such trade-off patterns have been used to infer which species are more vulnerable to climate change because they cannot evolve both high tolerance and plasticity (Heerwaarden & Kellermann, 2020). In this sense, P. ulyssiponensis may be at the greatest risk from future temperature increases and from the studied species is the most susceptible to climate change. This work provides information about the effect of temperature on the respiration rate in four species that occur on the rocky shores of mainland Portugal. It allows a better understanding of limpet species sensitivity to temperature increases, under climate change scenarios with possible extreme events, such as marine heatwaves. Extreme warm water conditions, as experienced during marine heatwaves, can exceed the thermal tolerance of marine taxa and lead to changes in the structure and function of intertidal communities (Wolfe et al., 2020; Amstutz et al., 2021). This research also provides new data about the thermal limits for three Patella spp. from mainland Portugal. It is a valuable step to improve our understanding of the biological responses of limpets in the context of climate change, with the associated consequences upon the algal assemblages, but also as a starting point for future studies.

Overall, the results of this research could be integrated in studies to foresee the potential impacts of climate change on limpet distributions and the potential of these species to persist or to adapt under different climate scenarios.

Acknowledgements. The authors would like to thank João Franco, for his availability to collect and send us one of the species used (from the central area of the country), which allowed us to carry out the experimental work. We are grateful to the two anonymous reviewers for their constructive comments on this manuscript.

Author contributions. DF: literature review, fieldwork, wrote the first draft of the manuscript, analysing the data, interpreting the findings. FA: designing the study, supervision, providing laboratory facilities and instruments, writing – review and editing. ISP: providing laboratory facilities and equipment, writing – review and editing. DB: formulating the research questions, designing the study, fieldwork, supervision, analysing the data, interpreting the findings, writing – review and editing. All authors contributed to the refinement of the final manuscript.

Financial support. This research was supported by the Portuguese project SEEINGSHORE (PTDC/BIA-BMA/31893/2017), co-financed by NORTE 2020, Portugal 2020 and the European Union through the ERDF, and by FCT through national funds. This research was partially supported by national funds through FCT – Foundation for Science and Technology within the scope of UIDB/04423/2020 and UIDP/04423/2020. This research has been additionally funded from the European Union's Horizon 2020 research and innovation programme under grant agreement No. 869300 'FutureMARES'. Débora Borges was additionally funded by the project BlueForesting (PT-INNOVATION-0077) financed by the EEA Grants.

Conflict of interest. The authors declare no conflict of interests.

References

- Allen AP, Gillooly JF and Brown JH (2005) Linking the global carbon cycle to individual metabolism. *Functional Ecology* **19**, 202–213.
- Amstutz A, Firth LB, Spicer JI and Hanley ME (2021) Facing up to climate change: community composition varies with aspect and surface temperature in the rocky intertidal. *Marine Environmental Research* 172, 105–482. doi: 10.1016/j.marenvres.2021.
- Blackmore DT (1969) Studies of Patella vulgata L.: growth, reproduction and zonal distribution. Journal of Experimental Marine Biology and Ecology 3, 200–213.
- Boaventura D, Ré P, Fonseca LD and Hawkins SJ (2002) Intertidal rocky shore communities of the continental Portuguese coast: analysis of distribution patterns. *Marine Ecology* **23**, 69–90.
- **Borges CDG, Doncaster CP, Maclean MA and Hawkins SJ** (2015) Broad-scale patterns of sex ratios in *Patella* spp.: a comparison of range edge and central range populations in the British Isles and Portugal. *Journal of the Marine Biological Association of the United Kingdom* **95**, 1141–1153.
- Branch GM and Newell RC (1978) A comparative study of metabolic energy expenditure in the limpets Patella cochlear, P. oculus and P. granularis. Marine Biology 49, 351–361.
- Branch GM, Borchers P, Brown CR and Donnelly D (1988) Temperature and food as factors influencing oxygen consumption of intertidal organisms, particularly limpets. *American Zoologist* 28, 137–146.
- Brown JH, Gillooly JF, Allen AP, Savage VM and West GB (2004) Toward a metabolic theory of ecology. *Ecology* 85, 1771–1789.
- Casal G, Aceña-Matarranz S, Fernández-Márquez D and Fernández N (2018) Distribution and abundance patterns of three coexisting species of *Patella (Mollusca gastropoda)* in the intertidal areas of the NW Iberian Peninsula: implications for management. *Fisheries Research* 198, 86–98.
- Chapperon C, Volkenborn N, Clavier J, Séité S, Seabra R and Lima FP (2016) Exposure to solar radiation drives organismal vulnerability to climate: evidence from an intertidal limpet. *Journal of Thermal Biology* 57, 92–100.
- Coleman RA, Underwood AJ, Benedetti-Cecchi L, Aberg P, Arenas F, Arrontes J, Castro J, Hartnoll RG, Jenkins SR, Paula J, Santina PD and Hawkins SJ (2006) A continental scale evaluation of the role of limpet grazing on rocky shores. *Oecologia* 147, 556–564.

- **Davies PS** (1966) Physiological ecology of *Patella*: I. The effect of body size and temperature on metabolic rate. *Journal of the Marine Biological Association of the United Kingdom* **46**, 647–658.
- **Davies PS** (1967) Physiological ecology of Patella: II. Effect of environmental acclimation on the metabolic rate. *Journal of the Marine Biological Association of the United Kingdom* **47**, 61–74.
- **Evans RG** (1948) The lethal temperatures of some common British littoral molluscs. *Journal of Animal Ecology* **17**, 165–173.
- Firth LB and Crowe TP (2010) Competition and habitat suitability: small-scale segregation underpins large-scale coexistence of key species on temperate rocky shores. *Oecologia* 162, 163–174.
- Folguera G, Bastías J, Caers JM, Rojas MD, Piulachs X, Bellés X and Bozinovic F (2011) An experimental test of the role of environmental temperature variability on molecular, physiological and life-history traits: implications for global warming. *Comparative Biochemistry and Physiology Part A: Molecular & Integrative Physiology* 159, 242–246.
- Gillooly JF, Brown JH, West GB, Savage VM and Charnov EL (2001) Effects of size and temperature on metabolic rate. *Science (New York, N.Y.)* 293, 1-6.
- Guerra MT and Gaudencio MJ (1986) Aspects of the ecology of *Patella* spp. on the Portuguese coast. *Hydrobiologia* **142**, 57–69.
- Harley CDG, Denny MW, Mach KJ and Miller LP (2009) Thermal stress and morphological adaptations in limpets. *Functional Ecology* 23, 292–301.
- Hawkins SJ, Sugden HE, Mieszkowska N, Moore PJ, Poloczansk E, Leaper R, Herbert RJH, Genner MJ, Moschella PS, Thompson RC, Jenkins SR, Southward AJ and Burrows MT (2009) Consequences of climate-driven biodiversity changes for ecosystem functioning of North European rocky shores. *Marine Ecology Process Series* 396, 245–259.
- Heerwaarden BV and Kellermann V (2020) Does plasticity trade off with basal heat tolerance? *Trends in Ecology & Evolution* 35, 874–885.
- Helmuth B, Mieszkowska N, Moore P and Hawkins SJ (2006) Living on the edge of two changing worlds: forecasting the responses of rocky intertidal ecosystems to climate change. *Annual Review of Ecology, Evolution and Systematics* 37, 373–404.
- Helmuth B, Yamane L, Lalwani S, Matzelle A, Tockstein A and Gao N (2011) Hidden signals of climate change in intertidal ecosystems: what (not) to expect when you are expecting. *Journal of Experimental Marine Biology and Ecology* **400**, 191–199.
- Hoegh-Guldberg O and Bruno JF (2010) The impact of climate change on the World's marine ecosystems. Science (New York, N.Y.) 328, 1523–1529.
- LathLean JA, Seuront L and Ng TPT (2017) On the edge: the use of infrared thermography in monitoring responses of intertidal organisms to heat stress. *Ecological Indicators* 81, 567–577.
- Lawler JJ (2009) Climate change adaptation strategies for resources management and conservation planning. Annals of the New York Academy of Sciences 1162, 79–98.
- Lima FP, Gomes F, Seabra R, Wethey DS and Seabra MI (2016) Loss of thermal refugia near equatorial range limits. *Global Change Biology* 22, 254–263.
- Lima FP, Queiroz N, Ribeiro PA, Hawkins SJ and Santos AM (2006) Recent changes in the distribution of a marine gastropod, *Patella rustica* Linnaeus, 1758, and their relationship to unusual climatic events. *Journal of Biogeography* 33, 812–822.
- Magris RA, Pressey RL, Weeks R and Ban NC (2014) Integrating connectivity and climate change into marine conservation planning. *Biological Conservation* 170, 207–221.
- Mertens NL, Russell BD and Connell SD (2015) Escaping herbivory: ocean warming as a refuge for primary producers where consumer metabolism and consumption cannot pursue. *Oecologia* 179, 1223–1229.
- Mieszkowska N, Burrows MT, Hawkins SJ and Sugden H (2021) Impacts of pervasive climate change and extreme events on rocky intertidal communities: evidence from long-term data. *Frontiers in Marine Science* 8, 642–764.
- Moore PJ, Thompson RC and Hawkins SJ (2011) Phenological changes in intertidal con-specific gastropods in response to climate warming. *Global Change Biology* **17**, 709–719.
- Newell RC and Branch GM (1980) The influence of temperature on the maintenance of metabolic energy balance in marine invertebrates. *Advances in Marine Biology* 17, 329–396.
- Newell RC and Roy A (1973) A statistical model relating the oxygen consumption of a mollusk (*Littorina littorea*) to activity, body size, and environmental conditions. *Comparative Physiology* **46**, 253–275.
- O'Connor MI (2009) Warming strengthens an herbivore-plant interaction. *Ecology* **90**, 388–398.

- O'Connor MI, Gilbert B and Brown CJ (2011) Theoretical predictions for how temperature affects the dynamics of interacting herbivores and plants. *American Naturalist* 178, 626–638.
- Pereira SG, Lima FP, Queiroz NC, Ribeiro PA and Santos AM (2006) Biogeographic patterns of intertidal macroinvertebrates and their association with macroalgae distribution along the Portuguese coast. *Hydrobiologia* 555, 185–192.
- **Prusina I, Balic-Ezgeta D, Ljubimir S, Dobroslavic T and Glamuzina B** (2014) On the reproduction of the Mediterranean keystone limpet *Patella rustica*: histological overview. *Journal of the Marine Biological Association of the United Kingdom* **94**, 1651–1660.
- Ribeiro PA, Xavier R, Santos AM and Hawkins SJ (2009) Reproductive cycles of four species of *Patella* (Mollusca: Gastropoda) on the northern and central Portuguese coast. *Journal of the Marine Biological Association of the United Kingdom* **89**, 1215–1221.
- Sagarin RD, Barry JP, Gilman SE and Baxter CH (1999) Climate-related change in an intertidal community over short and long time scales. *Ecological Monographs* 69, 465–490.
- Seabra R, Wethey DF, Santos AM, Gomes F and Lima FP (2016) Equatorial range limits of an intertidal ectotherm are more linked to water than air temperature. *Global Change Biology* 22, 3320–3331.
- Seabra R, Wethey DS, Santos AM and Lima FP (2011) Side matters: microhabitat influence on intertidal heat stress over a large geographical scale. *Journal of Experimental Marine Biology and Ecology* 400, 200–208.
- Silva A, Boaventura D and Ré P (2003) Population structure, recruitment and distribution patterns of *Patella depressa* Pennant, 1777 on the central Portuguese coast. *Boletin Instituto Espanol de Oceonografia* 19, 461–471.
- Somero GN (2002) Thermal physiology and vertical zonation of intertidal animals: optima, limits, and costs of living 1. *Integrative and Comparative Biology* 42, 780–789.
- Sousa LL, Seabra R, Wethey DS, Xavier R, Queiroz N, Zenboudji S and Lima FP (2012) Fate of a climate-driven colonisation: demography of newly established populations of the limpet *Patella rustica* Linnaeus, 1758, in northern Portugal. *Journal of Experimental Marine Biology and Ecology* 438, 68–75.
- Spalding MD, Fox HE, Allen GR, Davidson N, Ferdana ZA, Finlayson M, Halpern BS, Jorge MA, Lombana A, Lourie SA, Martin KD, McManus E, Molnar J, Recchia CA and Robertson J (2007) Marine ecoregions of the world: a bioregionalization of coastal and shelf areas. *Bioscience* 57, 573–583.
- Stillman JH (2019) Heat waves, the new normal: summertime temperature extremes will impact animals, ecosystems, and human communities. *Physiology* 34, 86–100.
- Tait RV and Dipper FA (1998) Elements of Marine Ecology, 4th edition. Oxford: Keyword Publishing.
- Tang J, Ikediala JN, Wang S, Hansen JD and Cavalieri RP (2000) High-temperature short time thermal quarantine methods. Postharvest Biology and Technology 21, 129–145.
- Underwood AJ (1979) The ecology of intertidal gastropods. Advance in Marine Biology 16, 111–210.
- **Underwood AJ and Jernakoff P** (1981) Effects of interactions between algae and grazing gastropods on the structure of a low-shore intertidal algal community. *Oecologia* **48**, 221–233.
- Vinagre C, Dias M, Cereja R, Abreu-Afonso F, Flores AAV and Mendonça V (2019) Upper thermal limits and warming safety margins of coastal marine species – indicator baseline for future reference. *Ecological Indicators* 102, 644–649.
- Vinagre C, Mendonça V, Cereja R, Abreu-Afonso F, Dias M, Mizrahi D and Flores AAV (2018) Ecological traps in shallow coastal waters – potential effect of heat-waves in tropical and temperate organisms. *PLoS ONE* 13, 1–17. doi: 10.1371/journal.pone.0192700.
- Watson SA, Morley SA, Bates AE, Clark MS, Day RW, Lamare M, Martin M, Southgate PC, Tan KS, Tyler PA and Peck LS (2014) Low global sensitivity of metabolic rate to temperature in calcified marine invertebrates. *Oecologia* 174, 45–54.
- Wolfe K, Nguyen HD, Davey M and Byrne M (2020) Characterizing biogeochemical fluctuations in a world of extreme: a synthesis for temperate intertidal habitats in the face of global change. *Global Change Biology* 26, 3858– 3879.
- Yvon-Durocher G, Jones JI, Trimmer M, Woodward G and Montoya JM (2010) Warming alters the metabolic balance of ecosystems. *Philosophical Transactions of the Royal Society* **365**, 2117–2126.